DEFINING WATERSHED-SCALE EVAPORATION USING A NORMALIZED DIFFERENCE VEGETATION INDEX

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ABSTRACT: Monthly composites of the Normalized Difference Vegetation Indices (NDVI), derived from the National Oceanic and Atmospheric Administration's (NOAA) Advanced Very High Resolution Radiometer (AVHRR), were transformed linearly into monthly evaporation rates and compared with detailed hydrologic-model simulation results for five watersheds across the United States. Model-simulated monthly evaporation values showed high correlations (mean $R^2 = .77$) with NDVI-derived evaporation estimates. These latter estimates, used in a classical water balance model, resulted in equally accurate simulations of monthly runoff than when the model was run to estimate monthly evaporation via soil moisture accounting. Comparison of NDVI-derived evaporation estimates with pan data showed promise for transforming NDVI values into evaporation estimates under both wet and water-limiting conditions without resorting to the application of any kind of calibrated hydrologic models.

(KEY TERMS: watershed evaporation; vegetation index.)

INTRODUCTION

Evapotranspiration is the loss of water from the Earth's surface in vapor form. It occurs as evaporation from open water and moist soil surfaces and as transpiration from living plants. Since the physical process is the same in either case (i.e., vaporization of water), the term "evaporation" is usually adequate to cover all processes of vaporization (Brutsaert, 1982:1). In this paper the term "evaporation" is used unless specified otherwise. Evaporation is a fundamental part of the hydrologic cycle since on a global scale two-thirds of the precipitation over land surfaces is soon lost to it (Brutsaert, 1982:4). Reliable evaporation estimates are critical to the fields of hydrology, meteorology and climatology (Parlange et al., 1995) because evaporation allows for the transfer of significant amounts of energy between the Earth's surface and its atmosphere. This efficient energy transfer is due to the large amount of latent heat involved in the vaporization of water, and as such, it has a great impact on the global circulation of the atmosphere and oceans and, consequently, on the Earth's climate (Luthi et al., 1997). Global Circulation Models (GCM) describing the evolution of weather and climate turned out to be quite sensitive to the hydrologic budgets of the continents (Committee on Opportunities in the Hydrologic Sciences, 1991).

The various equations for estimating areal evaporation can be expressed in the following general form: $E = \alpha PE$, where $E$ is actual and $PE$ is potential evaporation (i.e., evaporation rate under unlimited water availability), and $\alpha$ is the Bodyko-Thornthwaite-Mather parameter (Parlange et al., 1995). $\alpha$ is a function of soil moisture conditions and is unity until some measure of field capacity is reached, when its value decreases to zero with the drying of the soil (Parlange et al., 1995). While evaporation estimates have been proven effective over relatively homogeneous terrain (in terms of vegetation, relief, soil type), at a scale smaller than $10^4$ m (Crago and Brutsaert, 1992), their applicability at the watershed-scale (i.e., larger than $10^4$ m) is hindered by uncertainties in estimating effective, watershed-representative values of the parameters involved in formulating $E$, since the parameter $\alpha$ itself is generally dependent on the spatial distribution of the soil-moisture content and the soil-plant-atmosphere relationships (Parlange et al., 1995). One way to approach the problem of parameterization is the application of remotely sensed data.

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In recent years, there has been much progress in estimating the parameters involved in evaporation calculations using remote sensing techniques. Studies estimating atmospheric parameters (i.e., near-surface air temperature, wind and water-vapor gradient) include applications of such remote sensing instruments as: (a) sodars (Quintarelli, 1993; Kaimal and Finnigan, 1994; Thomas and Vogt, 1993a,b); (b) radars (e.g., Ralph et al., 1993); (c) and lidars (Galchen et al., 1992; Eichinger et al., 1993a,b, 1994; Parlange and Katul, 1995). While these techniques can be airborne or ground based, the remote sensing of general surface properties, such as surface temperature (e.g., Chen et al., 1997; Smith et al., 1997), surface soil moisture (e.g., Chen et al., 1997; Chauhan, 1997; Jackson et al., 1996), albedo (e.g., Lafleur et al., 1997), and vegetative cover (e.g., Wittich and Hansing, 1995), all affecting evaporation, may also rely on satellite-based sensors, allowing for data acquisition at a much larger scale.

A detailed description of the regulatory role of vegetation in the process of transpiration has also been the focus of extensive multi-disciplinary research (e.g., Bunce, 1997; Desborough, 1997; Granier et al., 1996; Koster and Milly, 1997; Prazak et al., 1996; Schreiber and Riederer, 1996). This is of special importance since transpiration makes up 90-95 percent of the total evaporation of vegetated surfaces (Maidment, 1993); thus defining transpiration rates provides a good estimate of evaporation over vegetated surfaces.

In the late 1970s, the application of remotely sensed vegetation indices for crop-yield monitoring and forecasting emerged (Tucker et al., 1979). By the mid 1980s, an especially useful combination of the spectral response of vegetation in different wavelengths had been introduced and was given the name of “Normalized Difference Vegetation Index” (NDVI). NDVI is a combination of the spectral response of vegetation in the near-infrared (NIR; i.e., .73-1.1 µm for AVHRR data) and red bands (R; i.e., .55-.68 µm for AVHRR data) (Tarpley et al., 1984)

\[
NDVI = \frac{NIR - R}{NIR + R}
\]

(1)

NDVI was demonstrated to be sensitive to changes in vegetation conditions since it is directly influenced by the chlorophyll’s absorption of the sun’s radiation (Tucker et al., 1985). Because the chlorophyll status integrates the effects of numerous environmental factors, NDVI has been empirically related to the following components of the hydrological cycle over a wide range of spatial and temporal scales: soil moisture (Walsh, 1987; Henricksen and Durkin, 1986; Choudhury and Golus, 1988; Farrar et al., 1994; Nicholson et al., 1996), precipitation (Tucker et al., 1985; Choudhury and Tucker, 1987; Seguin et al., 1989; Nicholson et al., 1990; Davenport and Nicholson, 1993; Schultz and Halpert, 1993; Di et al., 1994; Nicholson and Farrar, 1994; Grist et al., 1997; Yang et al., 1997), and evaporation (Running and Nemani, 1988; Kerr et al., 1989; Cihlar et al., 1991; Gao et al., 1992; Seevers and Ottmann, 1994; Nicholson et al., 1996; Szilagyi et al., 1998).

Running and Nemani (1988) regressed NOAA AVHRR-derived NDVI values against weekly forest evaporation rates calculated by an ecosystem model and found a linear relationship between the two. Kerr et al. (1989) applied NDVI values for the rainy season (June-September) of 1986 over Senegal to estimate evaporation, and found a strong linear relationship (R\(^2\) = .98) between cumulative NDVI and the 20-day shifted evaporation, estimated by a water-balance model. Cihlar et al. (1991), applying a different water-balance model in Canada, found a similarly strong linear relationship (R\(^2\) = .92) between the two variables, shifted by two weeks, using bi-weekly compilation periods for different soil type and vegetation combinations during the 1986 growing season (April-August). Gao et al. (1992) related six months of daily NDVI to the evaporative fraction (defined as the ratio of latent heat flux to the sum of latent and sensible heat fluxes) measured by fast response instruments at a single location during the First International Field Experiment (FIFE). Desjardins et al. (1992) related the greenness index (i.e., NIR/Red) of a 225 km\(^2\) area to latent heat fluxes measured by an aircraft for selected days, also during FIFE. Seevers and Ottmann (1994) used thematic mapper NDVI values to estimate evaporation of different irrigated crops on selected days. Szilagyi et al. (1998) demonstrated that the NDVI versus evaporation relationship on a monthly basis is at least moderate (R\(^2\) = .64) in a semi-arid environment as well, provided the hydrologic model-estimated evaporation data are shifted by one month later relative to the monthly mean NDVI values. Even the de-seasoned variables showed at least a weak correlation (R\(^2\) = .28) suggesting that the NDVI-versus-evaporation relationship may not solely be a consequence of similar seasonal cycles of the two variables.

These studies demonstrated the applicability of NDVI for estimating evaporation at a single location, over a field or a watershed. As Wiegand and Richardson (1990) argued, a strong relationship between NDVI and evaporation should not, in fact, be surprising because the green plant tissue, of which chlorophyll activity is measured by NDVI (Sellers, 1985), must be active both photosynthetically and transpirationally. Note that the majority of the authors cited above do not differentiate between evaporation and
transpiration in relation to NDVI. It should be pointed out, however, that NDVI can be physically related only to transpiration through Wiegand and Richardson's argument. The fact that NDVI relates to evaporation the same as it to transpiration is for reasons given earlier (i.e., transpiration is about 90 percent of the total evaporation of vegetated surfaces).

The above studies indicate a generally strong relationship between NDVI and evaporation, one that is linear or at least near-linear in nature. This means that NDVI values can be transformed into evaporation rates via a simple linear transformation. Below we show that NDVI-derived evaporation estimates result in simulated runoff of at least equal accuracy to classical water balance model estimates of both evaporation and runoff. This raises the question whether validating NDVI-derived evaporation values with hydrologic- or water balance-modeled estimates, as routinely done, is the best way to go, especially, when one can compare modeled runoff, obtained via the above-mentioned evaporation estimates (i.e. NDVI- and model-derived), with accurate observations.

HYDROLOGIC MODEL AND STUDY SITE DESCRIPTIONS

Two hydrologic models were used in this study. The first is a semi-distributed watershed model run on an hourly basis. For a detailed description of the model, see Szilagyi and Parlange (1999). The model divides the watershed into subcatchments according to stream-order and calculates simultaneous water balances on each subcatchment at an hourly increment. The model is computationally demanding and requires detailed descriptions of catchment geomorphology, stream network, soil types, land use/land cover types, and aquifer characteristics. Evaporation, \( E \) in the model is estimated by the following equation (Beven, 1991)

\[
E = PE \left(1 - \frac{V_i}{V_{i0}}\right).
\]

where \( V_i \) is the potential storage space of the soil. For a completely dry soil the maximum potential storage space \( V_{i0} \) can be estimated as

\[
V_{i0} = fRDA
\]

where \( f \) is the drainable porosity of the soil, \( RD \) is the rooting depth of the vegetation and \( A \) is the area of the specific soil type-vegetation combination. The potential evaporation \( (PE) \) is calculated as (Jensen and Haise, 1963)

\[
PE = 0.016742 \cdot R \cdot (0.014 \cdot (1.8 \cdot T + 32) - 0.37),
\]

where \( R \) is the incident solar radiation, and \( T \) is the mean monthly temperature (Celsius).

In a previous study the model has been applied in two catchments, Mahantango Creek in Pennsylvania and Winters' Run in Maryland, for characterizing subsurface contribution to runoff. In this study we will use the model-calculated hourly evaporation estimates for the above two catchments and aggregate them over the month, the time unit of our investigation of deriving evaporation using NDVI observations. Note that one cannot expect the vegetation to respond to changes in soil-moisture or atmospheric conditions overnight. A one-month time interval, however, may be large enough to let the vegetation adjust to changes in soil moisture and atmospheric conditions (Narasimha et al., 1993).

In addition to the watershed model, a classical water balance model was applied for five catchments across the U.S. The model is a modified version (Vorosmarty et al., 1989) of the classical Thornthwaite-Mather (1957) water-balance accounting procedure. Input data to the model include monthly sums of precipitation and radiation, monthly averages of temperature, dominant vegetation cover, and soil type of the catchment. At a monthly increment, the model makes predictions for potential evaporation, actual evaporation \( (E) \), soil-moisture content \( (SM) \), and runoff.

During months when precipitation \( (P) \) is in excess of \( PE \), soil moisture can increase up to a maximum field capacity \( (FC) \) determined by soil texture and rooting depths. During dry months, when precipitation is exceeded by \( PE \), soil moisture becomes a function of potential water loss. The relevant equations for soil-moisture calculations applied in the model are

\[
\frac{d(SM)}{dt} = P - PE, \quad \text{if} \ P > PE, \ SM < FC
\]

\[
\frac{d(SM)}{dt} = 0, \quad \text{if} \ P > PE, \ SM = FC
\]

\[
\frac{d(SM)}{dt} = -\beta \cdot SM \cdot (PE - P), \quad \text{if} \ P < PE,
\]

where \( \beta \) is the slope of the moisture-retention function. The value of \( \beta \) can be calculated by
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\[
\beta = \frac{\ln(FC)}{(1.1282 \cdot FC)^{1.2756}}
\]

which is an empirical formula for the slope of the moisture-retention function. This formula allows the retention function to behave differently for different soil types.

Once soil moisture is determined, evaporation is calculated. Following Thornthwaite and Mather (1957), \( E \) is set equal to \( PE \) in months, when precipitation is greater or equivalent to \( PE \). During these months it is assumed that precipitation satisfies the water demands of the vegetation. During dry months, when precipitation is less than \( PE \), the monthly sum of \( E \) is calculated as

\[
E = P - \frac{d(SM)}{dt}, \quad \text{if} \quad P < PE,
\]

During wet months, when field capacity is attained and the evaporation need of the vegetation is satisfied, the surplus water either seeps down into the soil to appear as groundwater recharge or runs off.

In addition to the above two watersheds (Mahantango Creek and Winters' Run), where both models were used, the catchments where the water-balance model was applied, are the Weeping Water, Nebraska, catchment and two watersheds in the Little River basin, Georgia, watersheds B and F (Figure 1). Watershed F is the headwater sub-basin of Watershed B in Georgia.

**Figure 1. Locations of the Selected Watersheds.**

Mahantango Creek in east-central Pennsylvania is a tributary of the Susquehanna River, which is located in the non-glaciated part of the North Appalachian Ridge and Valley Region. The watershed is characterized by long ridges of 300-400 m in elevation, alternating with broad valleys, 150-300 m in elevation. The geology of the Mahantango watershed can be described, from northwest to southeast, as folded Pennsylvanian sandstone and shale, Mississippian sandstone and shale, and Devonian sandstone, siltstone, and shale (NASA-EOS Report, 1995). The moderately weathered channery or stony loam soils, characteristic of the catchment, are thin with poorly developed horizons (ARS-USDA, 1976). The catchment area is 423 km². The catchment experiences a humid climate with an annual precipitation of 1,140 mm. The typical vegetation cover is mixed forest dominated by deciduous trees.

Winter's Run is a small, humid catchment (94 km²) in northeastern Maryland draining into the Chesapeake Bay 20 km south of the Susquehanna River. It is located in the Piedmont region, which is characterized by gently rolling hills of metamorphic and igneous rock types (Schmidt, 1993). The silty loam soils covering the catchment are thin with poorly developed horizons (Schmidt, 1993). Annual precipitation is 1,020 mm; the predominant land use is crop and pasture.

Catchments B and F, near the town of Tifton, Georgia, in the Little River basin, are experimental watersheds operated by the U.S. Department of Agriculture. The watersheds are representative of the Coastal Plain Province of the eastern United States, which extends from New England along the Atlantic coast to Texas (Williams, 1985). Catchments B and F have a drainage area of 334 and 114 km², respectively. About half of the drainage area of each catchment is covered with mixed forests, while the other half is dryland crops and pasture (Sheridan, 1997). The watersheds are covered with Quaternary sediments, poorly-sorted sands interbedded with partly indurated sandy claystones and clays that are underlain by limestones over the Hawthorn Formation, an aquiclude at a depth of 1-3 m (ARS-USDA, 1976). The soils are permeable, and the infiltration rates are high (Williams, 1985). The surface topography is relatively flat (Shirmohamaddi et al., 1986). Less than 2 percent of the annual precipitation of 1,200 mm is lost to deep percolation (Williams, 1985; Shirmohamaddi et al., 1986). Climate in the region is characterized as humid subtropical with long, warm summers and short, mild winters (Sheridan, 1997). Precipitation occurs almost exclusively as rainfall throughout the year (Sheridan, 1997).

The Weeping Water catchment in eastern Nebraska has a drainage area of 624 km². It lies within the
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Drift Hills region which is characterized by a rolling hill topography formed by glaciers in the Pleistocene. Loess in thick-to-thin deposits mantles the entire region, permitting moderate infiltration (CSD-UNL, 1998). The soils are deep, and moderately well drained. The climate of the area is typically continental with an average annual precipitation of about 800 mm, with the majority of the precipitation occurring in June. More than 80 percent of the watershed is cropland and pasture. Irrigation, unlike the rest of Nebraska, is negligible in the area, which is the most humid part of the state.

EVAPORATION ESTIMATION

For watersheds with varied vegetation types, catchment-scale evaporation rates under unlimited water supply conditions can be estimated by pan evaporation values modified by an appropriate coefficient (Brutsaert, 1982:214; Dunne and Leopold, 1978:127; Rodda et al., 1976:113; Viessman et al., 1989:101; Wilson, 1970:35-36). Published values of the pan coefficient for type A pans vary between .31 and 1.32 (Dunne and Leopold, 1978:101) and can be a function of the month. Evaporation under unlimited water supply becomes equal to potential evaporation (PE), which can be estimated by Equation (4). During wet months the pan coefficient-modified pan evaporation rate must equal PE. Figure 2 compares the two evaporation estimates during wet months in the Mahantango Creek, Pennsylvania, watershed. In dry months, neither the pan evaporation nor Equation (4) can estimate actual evaporation, since this latter one also becomes a function of soil moisture. A water balance/watershed model must be calibrated, usually against observed runoff, before meaningful E estimates can be obtained. Table 1 lists the explained variance ($R^2$) between monthly simulated and observed runoff values for each watershed using the calibrated water balance and the watershed model. The modeled runoff captures more than 80 percent of the variance in monthly runoff during the optimization period and about 70 percent in the verification period. The mean absolute error (MAE) is about 10 mm/month, which is about one-third of the mean standard deviation.

To check if spatially (over the watershed) averaged NDVI can be successfully converted into monthly evaporation estimates, the following has been done. The starting point coordinates of the NDVI versus evaporation relationship can be obtained by assigning an NDVI value to the assumed zero evaporation rate. Strictly speaking, watershed transpiration is very rarely zero, even during winter months, due to the widespread distribution of evergreen coniferous vegetation at practically any latitude. After applying a widely used transformation of adding one to each raw NDVI value of Equation (1) and multiplying it by 100 (Seevers and Ottmann, 1994; Di et al., 1994), this value theoretically is 100, since the range of the original NDVI values is [0; 1] (Price, 1990). Seevers and Ottmann (1994), using Thematic Mapper data, found this value to be 105 in practice, while we found a value of 110 for the watersheds that have been selected for analysis (see Table 1 and Figure 1) using pre-processed (Yang et al., 1997) monthly Maximum Value Composited 1-km resolution AVHRR data (USGS/EROS Data Center) for the period 1990-1997, except 1994, when NDVI values were not available for most of the year due to satellite problems. The two satellites use slightly different wavelengths to define NDVI (Seevers and Ottmann, 1994). The NDVI value of 110 also corresponds to the spatially averaged values of NDVI for the selected watersheds in January. This is because chlorophyll activity is reduced to close to zero during winter in the selected catchments.


Next, the slope of the NDVI versus evaporation relationship had to be defined. This was accomplished by systematically changing the slope value in the equation and running the water balance model with the so-derived evaporation estimates. The slope value was retained that resulted in the highest correlation value between observed and modeled runoff and in the lowest mean absolute error. Table 1 lists the optimized linear transformations.

<table>
<thead>
<tr>
<th>Water's Run</th>
<th>Mahantango</th>
<th>Watershed B</th>
<th>Watershed F</th>
<th>Weeping Water</th>
</tr>
</thead>
<tbody>
<tr>
<td>Area (km²)</td>
<td>94</td>
<td>423</td>
<td>334</td>
<td>114</td>
</tr>
<tr>
<td>$m_0 (\sigma_0)$ [mm]</td>
<td>41.85 (23.76)</td>
<td>39.67 (39.72)</td>
<td>32.81 (46.23)</td>
<td>36.31 (48.56)</td>
</tr>
<tr>
<td>$m_{aM} (\sigma_{aM})$ [mm]</td>
<td>42.48 (20.19)</td>
<td>51.20 (32.41)</td>
<td>31.92 (44.93)</td>
<td>35.93 (48.72)</td>
</tr>
<tr>
<td>MAEM [mm]</td>
<td>9.43</td>
<td>20.05</td>
<td>6.67</td>
<td>8.51</td>
</tr>
<tr>
<td>$R^2_{M}; &lt;R^2_M&gt; = .83$</td>
<td>.70 .76*</td>
<td>.67 .83*</td>
<td>.94</td>
<td>.94</td>
</tr>
<tr>
<td>$m_{aN} (\sigma_{aN})$ [mm]</td>
<td>41.91 (18.15)</td>
<td>40.01 (34.41)</td>
<td>33.03 (46.02)</td>
<td>36.12 (49.47)</td>
</tr>
<tr>
<td>MAEN [mm]</td>
<td>8.77</td>
<td>13.20</td>
<td>7.70</td>
<td>10.38</td>
</tr>
<tr>
<td>$R^2_{N}; &lt;R^2_N&gt; = .83$</td>
<td>.76</td>
<td>.77</td>
<td>.93</td>
<td>.90</td>
</tr>
</tbody>
</table>

Estimated E vs NDVI Relationship

E = 1.7NDVI-187

TABLE 1b. Model Performance Statistics, Verification Period.

<table>
<thead>
<tr>
<th>Winter's Run</th>
<th>Mahantango</th>
<th>Watershed B</th>
<th>Watershed F</th>
<th>Weeping Water</th>
</tr>
</thead>
<tbody>
<tr>
<td>$m_0 (\sigma_0)$ [mm]</td>
<td>53.15 (29.11)</td>
<td>48.54 (46.73)</td>
<td>19.26 (31.56)</td>
<td>21.61 (31.46)</td>
</tr>
<tr>
<td>$m_{aM} (\sigma_{aM})$ [mm]</td>
<td>55.11 (27.41)</td>
<td>50.17 (37.54)</td>
<td>16.12 (30.24)</td>
<td>17.57 (31.78)</td>
</tr>
<tr>
<td>MAEM [mm]</td>
<td>12.64</td>
<td>18.93</td>
<td>6.85</td>
<td>7.44</td>
</tr>
<tr>
<td>$R^2_{M}; &lt;R^2_M&gt; = .69$</td>
<td>.65 .70*</td>
<td>.67 .74*</td>
<td>.90</td>
<td>.87</td>
</tr>
<tr>
<td>$m_{aN} (\sigma_{aN})$ [mm]</td>
<td>54.31 (24.19)</td>
<td>45.31 (34.39)</td>
<td>18.82 (35.58)</td>
<td>20.97 (34.62)</td>
</tr>
<tr>
<td>MAEN [mm]</td>
<td>11.41</td>
<td>19.32</td>
<td>7.84</td>
<td>7.92</td>
</tr>
<tr>
<td>$R^2_{N}; &lt;R^2_N&gt; = .73$</td>
<td>.62</td>
<td>.66</td>
<td>.86</td>
<td>.87</td>
</tr>
<tr>
<td>$R^2_{X}; &lt;R^2_X&gt; &gt; .77$</td>
<td>.76 .81*</td>
<td>.78 .85*</td>
<td>.67</td>
<td>.63</td>
</tr>
</tbody>
</table>

KEY: $m$: mean monthly runoff; $\sigma$: standard deviation of monthly runoff; MAE: mean absolute error between monthly observed and simulated runoff; $R$: correlation coefficient; $E$: monthly evaporation [mm]; $< >$: average value, taken over the watersheds (excluding starred values); $< >^*$: average value, taken over the watersheds (including starred values).

KEY TO SUBSCRIPTS: $o$: observed value; $s$: model-simulated value; $M$: water balance-modeled runoff; $N$: water balance-modeled runoff using linearly transformed NDVI for monthly evaporation estimation instead of model simulated one; $X$: water balance model-simulated and NDVI-derived monthly evaporation estimates.

KEY TO SUPERSCRIPTS: $^*$: values obtained by using the watershed model outputs.
RESULTS AND DISCUSSION

The results of comparing model-calculated and NDVI-derived monthly evaporation estimates can be found in Table 1. The mean $R^2$ value taken over the five watersheds between the two variables is .77, which translates to a mean correlation coefficient of almost .9. In the least humid watershed, the Weeping Water catchment in Nebraska, the highest correlation resulted when the evaporation values calculated by the water-balance model were shifted one month later, compared to the NDVI-derived estimates, similar to what has been reported by Szilagyi et al. (1998) in the case of a semi-arid watershed in western Nebraska. Kerr et al. (1989) and Cihlar et al. (1991) found similar shifts in their NDVI and model-evaporation data. The existence of this shift pinpoints the important role vegetation plays in soil-moisture regulation with increasing aridity of the prevailing climate. Figure 3 displays the one-month-shifted model- and NDVI-derived monthly evaporation estimates. In Figure 4 the two variables are plotted against each other with the 1:1 line shown. Very similar graphs to Figures 3 and 4 can be obtained for the remaining watersheds with concurrent monthly data. Table 1 summarizes the results.

Figures 3 and 4 do not tell which evaporation estimate is closer to reality. In the lack of measurable watershed evaporation, NDVI-derived evaporation estimates have routinely been compared to model-calculated values (Kerr et al., 1989; Cihlar et al., 1991). Since these latter ones are prone to potentially large errors, such as the NDVI-derived evaporation values, comparing two variables with unknown errors does not seem to be the best way to proceed. To decide which one is a better representation of reality one could use the two estimates and check how a model, using the estimates, captures runoff, which can generally be measured. Table 1 shows (see the $R^2$ and MAE values) that the NDVI-derived evaporation estimates result in at least the same accuracy of runoff simulation as the model-generated evaporation estimates.

Insofar one had to rely on some kind of a hydrologic model to transform the watershed-averaged NDVI values into monthly watershed evaporation estimates through optimizing model simulated runoff. Since the resulting evaporation values are of about the same accuracy as the model-calculated evaporation estimates, there seems to be not much to be gained by using NDVI. The real advantage, however, of using NDVI comes around when lack of data prohibits any hydrologic model applications. Such may be the case when runoff data are missing or when irrigation is significant in the area of interest, but its level is unknown. This is typical of Nebraska, which is among the leading states in the U.S. in irrigated acreage and water volumes and where the law does not require farmers to document their agricultural water consumption. In such a situation, NDVI may still provide an alternative for estimating watershed evaporation, as described below.

It was mentioned earlier that watershed evaporation can be estimated by pan evaporation rates in non-water-stressed periods through the application of a pan coefficient. Figures 5a-e display the pan
Figure 5. Pan Coefficient-Modified Pan Evaporation Versus Watershed-Averaged NDVI in Wet Months. The thick line is the water-balance model-optimized NDVI-vs-evaporation relationship; the thin line is the linear least-square estimator going through (110, 0). (a) Weeping Water, Nebraska; (b) Watershed B, Little River, Georgia; (c) Watershed F, Little River, Georgia; (d) Mahantango Creek, Pennsylvania; (f) Winters’ Run, Maryland.
evaporation values, reduced by the most frequently chosen pan coefficient value of .8 (Dunne and Leopold, 1978:134) and plotted against NDVI values for months with abundant soil moisture (i.e., when precipitation was unexpectedly high in the previous month). The same graphs contain the water balance-optimized NDVI-evaporation lines (thick). As can be seen, the water balance model-optimized NDVI-evaporation equations are very close to the ones one can obtain by drawing a (thin) line from the origin [110, 0] of the graphs that minimizes the scatter between this line and the observed data points. It means that without resorting to any hydrologic model the sought-after NDVI-evaporation relationship can be determined by simply using NDVI and pan evaporation observations during wet months. As a consequence, watershed evaporation can be estimated using only precipitation, pan evaporation and NDVI records, even in cases when missing hydrological data prohibits hydrologic model applications.

In summary, NDVI-derived monthly watershed evaporation estimates were validated via the application of a water-balance model using the NDVI-derived evaporation values in the model and comparing the simulated runoff with observational data. The linear transformation of watershed-averaged monthly Maximum Value Composited NDVI values into monthly evaporation estimates proved to be at least as accurate as the classical evaporation estimation method of the Thornthwaite-Mather (1957) water balance accounting procedure in combination with the Jensen-Haise (1963) technique. A promising advantage of applying NDVI data for monthly areal evaporation assessment over hydrologic/water balance models is that the former does not have the typical data input requirement of a hydrologic/water balance model, which, in turn, means a possibly broader range of practical applicability.

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LITERATURE CITED


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